

## Biodiversity effects on yield and unsown species invasion in a temperate forage ecosystem

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- **Background and Aims** Current agricultural practices are based on growing monocultures or binary mixtures over large areas, with a resultant impoverishing effect on biodiversity at several trophic levels. The effects of increasing the biodiversity of a sward mixture on dry matter yield and unsown species invasion were studied.
- **Methods** A field experiment involving four grassland species [two grasses – perennial ryegrass (*Lolium perenne*) and cocksfoot (*Dactylis glomerata*) – and two legumes – red clover (*Trifolium pratense*) and white clover (*Trifolium repens*)], grown in monocultures and mixtures in accordance with a simplex design, was carried out. The legumes were included either as single varieties or as one of two broad genetic-base composites. The experiment was harvested three times a year over three years; dry matter yield and yield of unsown species were determined at each harvest. Yields of individual species and interactions between all species present were estimated through a statistical modelling approach.
- **Key Results** Species diversity produced a strong positive yield effect that resulted in transgressive over-yielding in the second and third years. Using broad genetic-base composites of the legumes had a small impact on yield and species interactions. Invasion by unsown species was strongly reduced by species diversity, but species identity was also important. Cocksfoot and white clover (with the exception of one broad genetic-base composite) reduced invasion, while red clover was the most invaded species.
- **Conclusions** The results show that it is possible to increase, and stabilize, the yield of a grassland crop and reduce invasion by unsown species by increasing its species diversity.

**Key words:** Cocksfoot, *Dactylis glomerata*, diversity effect, invasion, legumes, perennial ryegrass, *Lolium perenne*, red clover, *Trifolium pratense*, simplex design, statistical modelling, transgressive over-yielding, white clover, *T. repens*.

### INTRODUCTION

The loss of natural ecosystems to agriculture and other human activities is projected to have a large global impact on biodiversity in the future (Chapin *et al.*, 2000; Tilman *et al.*, 2002). Areas under intensive agriculture, in which applications of high doses of mineral fertilizers to monoculture crops are a prominent feature, have a negative impact on biodiversity in Europe (Reidsma *et al.*, 2006). However, maintaining acceptable levels of agricultural productivity will continue to be a high priority, mainly because of an increasing global population that needs to be fed, but also due to the need for economic sustainability of the individual farmer. Hence, there is an urgent need to develop agricultural practices that can deliver high yields with adequate forage quality, and simultaneously promote biodiversity. Biodiversity encompasses a broad spectrum of biotic scales (Hooper *et al.*, 2005), with components ranging from genetic diversity within populations through to the functional diversity of species present and up-scale to ecosystem diversity.

Perennial grasslands occupied around 40 % of the land used for agriculture by the European Union prior to its recent

enlargement (Rath and Peel, 2005). In productive temperate grassland systems grass monocultures are often used in order to simplify management options (e.g. easier to predict the optimal harvesting time, less variation in feeding quality). Seed use data for the EU in the 1990s showed that by far the most commonly sown grassland species was perennial ryegrass (Kley, 1995) and this remains the case today. However, large inputs of artificial nitrogen fertilizer are required to maximize the productivity of such monocultures, and this is increasingly regarded by policy makers and public opinion as being environmentally unacceptable. In addition to high yields, farmers may also value a stable yield level between years. Multi-species mixtures could provide greater and more stable primary production through temporal, spatial and resource niche complementarities (Sanderson *et al.*, 2004; van Ruijven and Berendse, 2005; Roscher *et al.*, 2008).

Although there is considerable evidence for a strong and persistent effect of diversity on yield in natural plant systems (Hooper *et al.*, 2005; Balvanera *et al.*, 2006; Cardinale *et al.*, 2007; Schmid *et al.*, 2002), few studies have explored biodiversity effects in a realistic agronomic context (but see Tracy and Sanderson, 2004a; Sanderson *et al.*, 2005, Picasso *et al.*, 2008). A positive diversity effect leading to

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transgressive over-yielding (mixture performing better than the best monoculture) was found in the first year's results from an agronomic study (in which some of the data of the present study were included) at 28 sites in Europe in which mixtures of four species (two legumes and two grasses) were compared with the monoculture performance of these species (Kirwan *et al.*, 2007). Genetic diversity might be expected to provide additional benefits through improved resistance to biotic (e.g. Finckh *et al.*, 1999; Cox *et al.*, 2004; Pilet *et al.*, 2006) and abiotic environmental stresses (Hajjar *et al.*, 2008), and enhanced niche complementarity (Hooper *et al.*, 2005). Further, it has been proposed that multi-species mixtures should be more resistant to invasion by unsown species than crops composed of fewer species, and this has been confirmed in some studies, e.g. Hector *et al.* (2001), Tracy *et al.* (2004), Tracy and Sanderson (2004b) Picasso *et al.* (2008). Although the evidence from a meta-analysis of 44 biodiversity experiments that manipulated plant species richness is that transgressive over-yielding is not, in general, of wide occurrence and requires several years to develop (Cardinale *et al.*, 2007), this conclusion may partly be due to methodological difficulties in its assessment (Schmid *et al.*, 2008). Persistence and an increase in a positive diversity effect in natural plant systems with time have been noted (Hooper and Dukes, 2004; van Ruijven and Berendse, 2005; Bullock *et al.*, 2007; Cardinale *et al.*, 2007; Fargione *et al.*, 2007) but whether the effect persists through time in perennial agronomic grassland systems is of considerable interest.

In this paper results are reported from a field study comprising multi-species mixtures of four grassland species [two grasses and two legumes: perennial ryegrass (*Lolium perenne*) and cocksfoot (*Dactylis glomerata*) and red clover (*Trifolium pratense*) and white clover (*T. repens*)] using the approach described in Kirwan *et al.* (2007). The experiment also included a treatment in which the impact of intraspecific variation in the legume species was assessed by the use of single varieties versus broad genetic-base composites, comprising mechanical mixtures of seed of many varieties (Collins *et al.*, 2004). It was hypothesized that: (1) positive interactions between species would result in larger yields from mixtures than expected from the performance of the individual species sown as monocultures; (2) these benefits would persist over several years; (3) a broad genetic-base composite would provide a platform for further and sustained positive effects of mixing species; and (4) increased species diversity and the use of broad genetic-base composites would help to prevent invasion of unsown species into the swards.

## MATERIALS AND METHODS

### Site data and plant material

The field experiment was conducted at Svalöv in southern Sweden (56°55'N, 13°07' E, 55 m a.s.l.). The climate is cold-temperate with an annual mean temperature of 7.7 °C and an annual precipitation averaging 700 mm. The soil at the site was a sandy clay loam with a pH of 6.9 and containing 2.12 % organic matter, 1.3 g nitrogen, 0.36 cmol potassium and 44 mg phosphorus kg<sup>-1</sup> soil. Plots were established in late June 2003 by drilling, and initial invasion by unsown species

was controlled by spraying with the herbicide bentazon (which is tolerated by the species sown) shortly after the initial emergence of the crop. The species used were two grasses and two legumes, of which one species within each functional group was relatively fast establishing but not persistent (*Lolium perenne* L. and *Trifolium pratense* L.), and the other was relatively slow establishing and persistent (*Dactylis glomerata* L. and *Trifolium repens* L.). Information on the plant material is presented in Table 1. For each legume species there was one single variety of species (denoted N) and two broad genetic-base composites, comprising 'Broad Central European' (BC) and 'Broad Northern European' (BN) populations. The single varieties of species used (N) had been bred for agricultural use within the experimental area. The red clover BC composite contained seed of 11 commercial varieties bred for use in temperate regions, plus a small amount (23 % by weight) of a 'northern red clover composite' population from Norway/Denmark, and encompassed all categories employed by red clover breeders (early/late flowering; diploid/tetraploid). The red clover BN composite contained seed from a population which was created by the inter-crossing for two generations of 238 accessions comprising the complete red clover collection held in the Nordic Gene Bank, plus varieties from other parts of the world (mainly from the former USSR; Á. Helgadóttir, Agricultural University of Iceland, Reykjavik, pers. comm.). The white clover BC composite contained seed of 13 commercial varieties bred for use in temperate regions, plus a small amount (5 % by weight) of 'gene pool' material from Central Asia. The white clover BN composite contained seed from nine varieties bred for use in northern Scandinavia, plus ten commercial varieties bred for use in southern Scandinavia (P. Marum, Graminor AS, IIseng, Norway, pers. comm.).

Seeding rates of the respective species in monoculture in the high density treatment (see below) are found in Table 1. These corresponded to the rates used in official Swedish variety trials. Low density plots were sown at 60 % of these rates. In mixtures, the seeding rate of each species was a proportion (see below) of the monoculture seeding rates.

The experimental layout consisted of 66 communities. Thirty communities followed the simplex design described in Kirwan *et al.* (2007) with two grass and two legume (N)

TABLE 1. Plant material and monoculture seeding rates

Species	Population	Monoculture seeding rates (kg ha <sup>-1</sup> ) for high sowing density
Perennial ryegrass ( <i>Lolium perenne</i> )	Helmer	30
Cocksfoot ( <i>Dactylis glomerata</i> )	Dactus	16
Red clover ( <i>Trifolium pratense</i> )	Fanny, N	15
Red clover	Central European composite, BC	15
Red clover	Northern European composite, BN	15
White clover ( <i>Trifolium repens</i> )	Ramona, N	8
White clover	Central European, BC	8
White clover	Northern European, BN	8

monocultures and 11 mixtures of these four species sown at two densities. The 11 mixtures consisted of four mixtures dominated in turn by each species (sown proportions were 70 % of dominant and 10 % of each other species), six mixtures dominated in turn by pairs of species (40 % of each of the two species and 10 % of the other two) and the centroid community (25 % of each species). Eighteen communities had four monocultures [two grasses (same as before) and two legumes (BC)] at two densities and five mixtures of these four species sown at two densities (sown proportions were 70 % of the dominant and 10 % of each of the other species, plus the centroid community). Another 18 communities followed the same design but with BC replaced by BN. Experimental communities were randomly assigned to 8.8 m<sup>2</sup> plots. Data from one plot were omitted since an erroneous species mixture was sown to this plot. General information on simplex designs is available in Cornell (2002).

No harvests were taken in the establishment year. In the following three years the whole plots were harvested with an experimental plot harvester (Haldrup) 3 times a year, with a stubble height of approx. 5 cm (normal practice within this climatic zone). At each harvest biomass yield was determined and samples for analyses of dry matter (DM) and botanical composition were taken from all plots. White clover ingress into neighbouring plots was prevented by regular spraying of the borders with glyphosate. No fertilizer nitrogen was added to the experiment. In the final year of the experiment notes were taken of the unsown species identity present in all plots.

#### Statistical analysis

Annual sown DM yield and unsown species DM % for each year and sown DM yield for each harvest were analysed. The modelling approach of Kirwan *et al.* (2007) was used to relate sown yield ( $DM_{\text{yield}}$ ) for a single harvest or year to variables representing species sown proportions of monoculture seeding rates, overall density and species interaction. Sown proportions were denoted in the model as  $G_1$  (ryegrass) and

$G_2$  (cocksfoot) for the two grasses and  $L_{1N}$ ,  $L_{1BC}$  and  $L_{1BN}$  for the N, BC and BN genetic treatments, respectively, for the first legume species ( $L_1$ , red clover). A similar notation was used for the second legume species ( $L_2$ , white clover).  $DENS$  is a variable indicating high or low seeding density, coded  $-1$  for low and  $1$  for high. A single variable (evenness) was used to describe interspecific interaction in a mixture. The evenness ( $E$ ) of a mixture as defined in Kirwan *et al.* (2007) lies between 0 for a monoculture and 1 for a mixture in which the sown proportions of all species are equal (the centroid). They used  $E$  and  $E^2$  as explanatory variables to capture the effect of species interactions. The model fitted was

$$\begin{aligned} DM_{\text{yield}} = & \beta_1 G_1 + \beta_2 G_2 + \beta_3 L_{1N} + \beta_4 L_{1BC} + \beta_5 L_{1BN} + \beta_6 L_{2N} \\ & + \beta_7 L_{2BC} + \beta_8 L_{2BN} + \beta_9 DENS + \beta_{10} E_N + \beta_{11} E_{BC} + \beta_{12} E_{BN} \\ & + \beta_{13} E^2 + \varepsilon \end{aligned} \quad (1)$$

The coefficients  $\beta_1$  and  $\beta_2$  represent the monoculture responses of the grass species and the coefficients  $\beta_3$  to  $\beta_8$  represent the monoculture responses of the legume species, at average density. For example, 3.84, the estimate of  $\beta_1$  for year 1 (see Table S1 in Supplementary data available online), estimates the monoculture yield of perennial ryegrass for that year. The evenness coefficients ( $\beta_{10}$ ,  $\beta_{11}$  and  $\beta_{12}$ ) measure the effects of evenness for mixtures involving each of the legume populations.  $\beta_{13}$  is a quadratic effect of evenness that does not change with legume population. Thus, for mixtures involving N legumes, the species diversity effect (the contribution of interaction effects to yield of a mixture at evenness  $E_N$  over what could be expected from a proportional mixture of monoculture yields) is  $\beta_{10} E_N + \beta_{13} E_N^2$  estimated by  $5.62 E_N - 2.14 E_N^2$  (Table S1). For the centroid (equal sown proportion of all species),  $E_N = 1$  and the predicted diversity effect is  $5.62 - 2.14 = 3.48$  (Table 2). Finally,  $\varepsilon$  is a random error term assumed to be normally and independently distributed with

TABLE 2. Mean estimated species diversity effects for sown dry matter yield at the centroid ( $t \text{ ha}^{-1}$ ) of the three legume population treatments for each of the three harvest years with tests of significance; also shown for each year are yields ( $t \text{ ha}^{-1}$ ) at the centroid predicted both from monoculture performance only, and including the species diversity effect

	Harvest year 1			Harvest year 2			Harvest year 3		
	Estimate	s.e.	P	Estimate	s.e.	P	Estimate	s.e.	P
Estimated species diversity effect at the centroid									
N	3.48	0.277	<0.001	4.82	0.231	<0.001	3.7	0.241	<0.001
BC	3.96	0.339	<0.001	5.3	0.303	<0.001	4.18	0.311	<0.001
BN	3.21	0.349	<0.001	4.55	0.313	<0.001	3.44	0.321	<0.001
Estimated yield at the centroid based on monoculture performances									
N	7.14	0.173		3.81	0.142		1.99	0.149	
BC	7.04	0.175		4.05	0.143		2.16	0.150	
BN	7.34	0.193		4.04	0.160		1.78	0.168	
Estimated yield at the centroid including the species diversity effect									
N	10.62	0.220		8.63	0.180		5.70	0.189	
BC	11.00	0.306		9.35	0.266		6.34	0.275	
BN	10.56	0.308		8.59	0.267		5.22	0.276	

Comparisons among estimates within and across years are presented in the text. N denotes a legume variety, while BC and BN denote Broad Central European and Broad Northern European genetic-base composites, respectively.

mean of zero and constant variance. Differences were tested between coefficients for grass species ( $\beta_1$  vs.  $\beta_2$ ), for legume species ( $\beta_3$  vs.  $\beta_6$ ,  $\beta_4$  vs.  $\beta_7$ , and  $\beta_5$  vs.  $\beta_8$ ) and legume populations ( $\beta_3$  vs.  $\beta_4$  vs.  $\beta_5$  and  $\beta_6$  vs.  $\beta_7$  vs.  $\beta_8$ ), and the evenness effect was compared between different legume populations ( $\beta_{10}$  vs.  $\beta_{11}$  vs.  $\beta_{12}$ ) and they were tested against zero.

Repeated measures analyses were carried out using the MIXED procedure and univariate analyses using the GLM Procedure in the SAS/STAT software, Version 9.1 of the SAS System for Windows.

Repeated measures analyses of sown species biomass were performed across harvests within each year and for annual sown yields across the 3 years of the experiment based on model (1) augmented with terms for interactions with time, which were significant. An unstructured covariance matrix was used to describe the time-dependence among harvests or years (Verbeke and Molenbergh, 2000). The effects of year and legume population were tested on the estimated species diversity effect at the centroid and the average monoculture yield across years. Transgressive over-yielding was tested for each legume population using a non-parametric test similar to that described in Kirwan *et al.* (2007).

The percentage of total annual unsown species biomass in a repeated measures analysis across years was analysed using model (1) and including year effects as just described. Responses were log-transformed before analysis to reduce heteroscedasticity. Many mixed plots were free from unsown species in the two first harvest years (14 in each year), and the recorded zero values from these plots were replaced by 0.0014, the lowest recorded non-zero value, before transformation.

There was no interaction between density and the other model variables. Model (1) (and its extensions for repeated measures analysis across years) generally emerged as either

the dominant or the best of a number of alternative models for overall sown yield, for most harvests and years and for log of unsown species yield (by dominant it is meant that the explanatory power of the evenness-based species diversity effects greatly exceeds any refinement introduced by partitioning the evenness effect into more coefficients as described below). Alternative models included a model with separate interaction coefficients for each pairwise interaction, and a model that included only two pairwise interaction coefficients, one for an interaction between any grass and any legume and another for grass–grass or legume–legume interaction (Kirwan *et al.*, 2007). Other models allowed the species diversity effect to vary with legume species. There was some evidence in favour of separate pairwise coefficients for each interaction for sown yield at some harvests and for a separate coefficient  $\times$  year interaction for the analysis of sown annual yield. In the latter analysis there was no clear pattern among the six interaction coefficients across years. Mixtures were not replicated and the variation around the regression model was used to estimate the residual mean square and the standard errors of coefficient estimates (Cornell, 2002).

## RESULTS

### Species-specific effects

Estimates of coefficients of model (1) fitted to annual sown yield as a repeated measure over 3 years show that there was a considerable and significant decline in monoculture yield for all species across years ( $P < 0.05$  for each test; Fig. 1, and Table S1 in Supplementary data, available online). The estimated DM yield of the cocksfoot monoculture was consistently greater than that of perennial ryegrass ( $P < 0.01$  in each

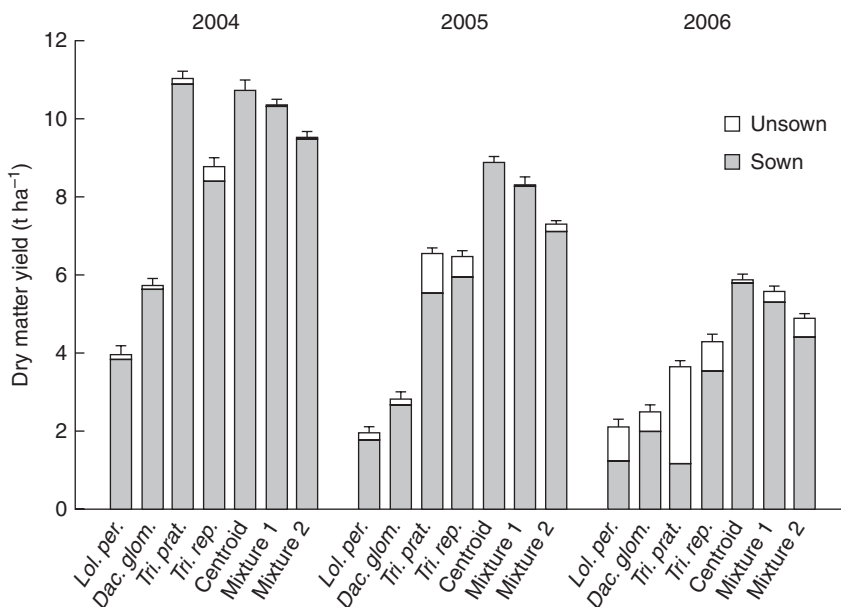


FIG. 1. Total yield predicted for each of the four monocultures and for three mixtures divided into sown and unsown species for each year. The four monocultures are *Dactylis glomerata*, *Lolium perenne*, *Trifolium pratense* (averaged over N, BC and BN) and *Trifolium repens* (averaged over N, BC and BN). The three mixture communities are the Centroid (0.25, 0.25, 0.25, 0.25), Mixture 1 (the average of the mixtures with two species dominant, e.g. 0.4, 0.4, 0.1, 0.1) and Mixture 2 (the average of the mixtures with one species dominant, e.g. 0.7-0.1 0.1 0.1). Standard error bars are given for the predicted total yield.



harvest year; Table S1 in Supplementary data, available online). White clover consistently outyielded the two grasses over the 3 years ( $P < 0.01$  for each test except  $L_{2BN}$  vs.  $G_2$  in year 3), while red clover did so in the first 2 years ( $P < 0.001$  for each test). There was generally no strong or consistent effect of legume populations in the monoculture performances of either legume species for any year. Red clover outyielded white clover in the first year ( $P < 0.001$  for N and BC populations,  $P = 0.06$  for BN) but underyielded it in the third year ( $P < 0.01$  for all populations). Density never affected DM yield.

While the cumulative yield for each year was little affected by the broad genetic-base composite treatment, the different legume populations nonetheless showed a different yielding pattern within each year (see Table S2 in Supplementary data, available online). Monocultures of the BC red clover populations had higher estimated yield ( $\beta_4$  vs.  $\beta_3$  and  $\beta_4$  vs.  $\beta_5$  red clover) than the N and BN red clover populations in the final harvest of each year, significantly so ( $P < 0.01$ ) in the first 2 years. Further, monocultures of the BC white clover populations had higher estimated yield than the BN white clover populations in the final harvest of each year ( $\beta_7$  vs.  $\beta_8$  white clover), significantly so in 2 of the 3 years. In the other harvests differences were rarely significant.

#### Diversity effects

Evidence of transgressive over-yielding was found in 3-year accumulated yield in the mixtures using N and BC legumes ( $P < 0.05$ ; Table 3). Transgressive over-yielding was observed in the second and the third years for mixtures using all legume populations ( $P < 0.05$ ).

There was a strong and persistent species diversity effect irrespective of the legume population used. The linear evenness effects were always positive and greater in magnitude than the quadratic evenness effects (Table S1 in Supplementary data available online), giving a positive species diversity effect that remained strong over the 3 years (Table 2). The linear evenness effect varied with both legume population used ( $P < 0.001$ ) and year ( $P < 0.001$ ), and the quadratic evenness effect varied with year ( $P < 0.001$ ; Table S1 in Supplementary data, available online). The maximum species diversity effect occurred at a level of evenness at or close to 1, the centroid. The species diversity effect was calculated at the centroid for each genetic diversity level  $\times$  year combination and also the mean yield calculated at the centroid predicted from the monoculture yields of species, i.e. with no diversity effect added, and the mean yield including the species diversity effect (Table 2). The species diversity effect at the centroid increased between the first and second years ( $P < 0.001$ ; Table 2) and in the third year declined to levels similar to those in the first year ( $P < 0.001$ ; Table 2) for each of them. Yield at the centroid predicted from an equal mixture of monoculture yields declined strongly over time, irrespective of the legume population used ( $P < 0.001$  for each test; Table 2). Predicted yield at the centroid including the species diversity effect also declined over time ( $P < 0.001$  for each test; Table 2) but far less rapidly than when based solely on monoculture performance. Thus, the species diversity effect on yield relative to yield predicted from the

TABLE 3. Results of tests for transgressive over-yielding

Year	Harvest	N	BC	BN
1	1	4	2	3
	2	0	2	0
	3	7	1	2
2	1	<b>11</b>	<b>5</b>	<b>5</b>
	2	<b>10</b>	4	0
	3	5	2	4
3	1	<b>11</b>	<b>5</b>	<b>5</b>
	2	<b>9</b>	2	1
	3	<b>11</b>	4	2
1		6	1	2
2		<b>11</b>	<b>5</b>	<b>5</b>
3		<b>11</b>	<b>5</b>	<b>5</b>
	1	<b>11</b>	<b>5</b>	<b>5</b>
	2	<b>8</b>	3	1
	3	<b>8</b>	3	4
Total for 3 years		<b>10</b>	<b>5</b>	3

The test is carried out for each of the three populations (N, BC and BN), for each harvest by year combination, for yield accumulated each year and for yield accumulated over all 3 years. N denotes a legume variety, while BC and BN denote Broad Central European and Broad Northern European genetic-base composites, respectively. The value given in the table is the test statistic\*,  $d$ , which is the absolute difference between the number of mixture means better than the best monoculture and worse than the worst monoculture. Bold indicates significance of the test ( $P < 0.05$ ). The test is based on the non-parametric test described in Kirwan *et al.* (2007).

\* Transgressive over-yielding occurs when the yield of a mixture community exceeds that of the highest-yielding monoculture of its component species. Here, a permutation test is used to determine the significance of transgressive over-yielding. There were 11 (4), 5 (4) and 5 (4) mixture communities (monoculture communities) in the three legume populations N, BC and BN, respectively. The average across replicates for each monoculture and mixture community was calculated within each legume population except for grass monocultures which were averaged across the legume populations (there was no difference between the grass monocultures for each legume population), giving 15, 9 and 9 community values for N, BC and BN respectively. The test statistic,  $d$ , is calculated from these community values as the absolute difference between the number of mixtures greater than the best monoculture and lower than the worst monoculture. Significance was tested using a permutation test that compares the result against that which would be expected by selecting 11 (or 5) values at random from 15 (or 9) and comparing them with the best and worst of the remaining 4 (or 4) values. The test is a two-sided test as the mixtures could be better than the best monoculture or worse than the worst monoculture (under-yielding) and the direction cannot be prejudged in advance. For more details see supplementary material in Kirwan *et al.* (2007).

respective monoculture performances was much higher in the second and third years than in the first year. Mixtures with the BC legumes produced a generally stronger species diversity effect than those with BN legumes ( $P < 0.05$ ), while the species diversity effect of the mixtures including the N legume was not significantly different from that of either of the two broad genetic-base composites (Table 2).

#### Invasion by unsown species

The effect of species diversity was very marked and persisted through time (significant species diversity effect each year  $P < 0.001$ ), and the predicted level of unsown species at the centroid was  $< 2\%$  even in the third year (Table 4).

The most striking monoculture effects observed were the ability of cocksfoot and white clover (N and BC) to resist invasions by unsown species (Table 4, and Table S3 in

TABLE 4. Estimated biomass percentage of unsown species in the respective monocultures and at the centroid for each of three harvest years (HY)

Parameter	Estimated weed content (%)		
	HY 1	HY 2	HY 3
$\beta_1(G_1 - \text{ryegrass})$	1.5	3.4	41.3
$\beta_2(G_2 - \text{cocksfoot})$	0.2	0.7	5.9
$\beta_3(L_{1N} - \text{red clover})$	1.2	12.6	68.2
$\beta_4(L_{1BC} - \text{red clover})$	1.2	12.6	68.2
$\beta_5(L_{1BN} - \text{red clover})$	1.0	10.6	57.2
$\beta_6(L_{2N} - \text{white clover})$	1.4	2.3	6.0
$\beta_7(L_{2BC} - \text{white clover})$	1.3	2.2	5.8
$\beta_8(L_{2BN} - \text{white clover})$	5.9	10.1	26.2
Centroid			
N	0.1	0.1	1.1
BC	0.1	0.1	1.1
BN	0.1	0.1	1.5

N denotes legume varieties, while BC and BN denote Broad Central European and Broad Northern European genetic-base composites, respectively. The centroid is the mixture sown with equal proportions of the four species.

Supplementary data available online). Even in the third year levels of unsown species were <6% in these monocultures, whereas they increased over years to high levels for monocultures of both ryegrass (41%) and red clover (around 60%). Using broad genetic-base composites appeared to have little effect on the amount of unsown species in red clover swards, but the BN white clover population contained a larger amount than the N or the BC populations ( $P < 0.01$ ; Table 4).

## DISCUSSION

The present study differs from many earlier biodiversity studies (e.g. Hector *et al.*, 1999; Tilman *et al.*, 2001; Roscher *et al.*, 2003) with respect to the species used (all of major agronomic importance in temperate regions) and the management carried out (aimed at maximizing biomass harvested; see Hooper *et al.*, 2005). The design of the experiment also differed from the above-mentioned studies in establishing a gradient of evenness with all species present in all mixed communities, instead of just increasing species richness as a treatment.

With respect to the four main hypotheses addressed in this study it was found that species diversity had a strong, persistent and positive effect on yield and on the ability of the sward to resist invasion by unsown species. However, using a broad genetic-base composite in two of the four species in the experiment had little impact on these ecosystem properties.

### Diversity effects on yield

It was hypothesized that the mixing of species would result in larger yields from the mixtures than would be expected from the individual species monoculture yields. That mixing of grasses with legumes, with a resulting niche complementarity for N (Palmborg *et al.*, 2005; Kahmen *et al.*, 2006), results in a positive yield effect has been shown in other studies (e.g. Spehn *et al.*, 2002). However, when a model that partitioned

the diversity effect into a term representing legume  $\times$  grass interaction was fitted, and a term jointly representing grass  $\times$  grass and legume  $\times$  legume interactions (Kirwan *et al.*, 2007) the two coefficients did not differ significantly, suggesting that the evenness model was a more parsimonious description of these data. This suggests that the effects of mixing grasses or mixing legumes had the same impact on increasing yield as did functional group mixing between grasses and legumes. Similar results have been reported by van Ruijven and Berendse (2003, 2005) and Hector *et al.* (2007). Thus, it appears that the characteristics of the species within each functional group were either of a complementary or a facilitating nature (Hooper *et al.*, 2005). The species diversity effect based on between-species interactions was always positive. In the centroid mixture the average contribution to yield of the species diversity effect were 3.55, 4.89 and 3.77 tons DM ha<sup>-1</sup> (+33%, +55% and +65% yield increase over mean monoculture performances) in the first, second and third years respectively. This diversity effect is stronger than the mean effect observed for the first harvest year across 28 European sites (Kirwan *et al.*, 2007).

The sustained species diversity effect and the declining monoculture effects suggest that the importance of this diversity contribution to yield increased with time. This was reflected in the increasing evidence of transgressive over-yielding for all legume populations as time advanced. Reports of transgressive over-yielding in biodiversity experiments carried out in the field with grassland species have been rare (e.g. Hooper and Dukes, 2004; Cardinale *et al.*, 2007; Lanta and Lepš, 2007; Schmid *et al.*, 2008; but see Schmid *et al.*, 2002). It is argued (Schmid *et al.*, 2008) that a wide range in monoculture yields makes it difficult to observe transgressive over-yielding. However, in some studies where transgressive over-yielding has been observed, the species involved have been high-yielding (Jolliffe and Wanjanu, 1999; Roscher *et al.*, 2005), the soil has been nutrient-rich (Roscher *et al.*, 2008) and species richness has been low (Roscher *et al.*, 2005), which was also the case in the present experiment. This suggests that transgressive over-yielding should be more common in an agricultural context, where soil fertility usually is good and the species used are selected on the basis of their yielding properties, than it is in natural communities in which low-yielding species frequently occur. Indeed, in an analysis of 44 plant experiments (39 grassland) with species number ranging from 6 to 32 (Cardinale *et al.*, 2007) the yield of the highest yielding monoculture was 93% above the yield of the average monoculture and transgressive over-yielding was only observed in 12% of cases. However, in the present data, the monoculture variation was lower than this but still of a substantial magnitude (the highest yielding monoculture exceeded the average monoculture by 51, 51 and 77% in harvest years 1, 2 and 3, respectively). So, the emergence of transgressive over-yielding over time here is not due to a relatively small species diversity effect being compared with rather similar yielding species in monoculture, but rather to the decline in monoculture yields over time and the persistence of a relatively large species diversity effect.

The persistent and increasing effect of species diversity on yield has important practical implications. In farming practice,

the application of fertilizer nitrogen by European grassland farmers is a common way to increase and/or maintain yield over time. Given the need to reduce the load of nitrogen into the European environment (e.g. Rougoor and van der Weijden, 2001) the practical benefit of this diversity effect is obvious, and the use of mixtures could therefore be recommended as a way of reducing inputs. It should be noted that no fertilizer nitrogen was added to this experiment.

Using broad genetic-base composite populations of the two legumes had a relatively small impact on yield in the current experiment, contrary to hypothesis (3). This could be due to the fact that traits are more likely to be similar within species than between different species. Similar results have been obtained in studies with white clover by Annicchiarico and Piano (1997) and Williams *et al.* (2003). Legumes, and red clover in particular, are susceptible to a vast range of pathogens that destroy sensitive genotypes (Frame *et al.*, 1998), and the loss of red clover plants due to diseases was very high in this experiment. Although there are no data on disease susceptibility of the red clover populations or pathogen load in the plots, the fact that yields of all red clover populations were similar in the final year suggest no difference in susceptibility to pest and diseases among them. Resistance of perennial plant species to abiotic stress during winter is often gained at the expense of late-season growth (e.g. white clover; Eagles and Othman, 1989). However, in the environment of southern Sweden the observed differences in end-of-season growth pattern of the legumes did not appear to have any impact on the overwintering abilities of the populations (as judged by monoculture spring yields). Attempts in the USA to improve the yield of a lucerne (*Medicago sativa*) sward by mixing varieties with a different seasonal growth pattern were short-lived, and this strategy was not recommended as a way to improve yield (Brummer *et al.*, 2002). Another explanation for the weak effect of the broad genetic-base composites might be the occurrence of selection of the most adapted individuals for the site and the management over time (e.g. Frankow-Lindberg, 1999, 2001). Joshi *et al.* (2001) found that red clover ecotypes performed best in their original sites compared with more distant sites, and their performance was poorer with increased transplanting distance. The broad genetic-base composites might therefore not have been well adapted to Swedish conditions. Hypothesis (3) was thus not confirmed, but it is possible that the use of a broad genetic-base composites of varieties specifically adapted to southern Sweden might have resulted in a more long-lasting complementary effect within the plant community.

#### *Diversity effects on unsown species invasion*

The species diversity effect on unsown species was very strong, confirming hypothesis (4). The nature of the relationship between species diversity and resistance to invasion by unsown species is a hotly debated topic among ecologists (e.g. Fridley *et al.*, 2007). In this experiment, where no weeding was carried out, both species diversity and species identity were important in controlling invasion. The main unsown species recorded was *Taraxacum officinale*, which is a wind-dispersed species that does not form a persistent soil

seed bank. This shows that the soil seed bank was a factor of little importance for invasion by unsown species at this site, and that any effect of the herbicide application at the very beginning of the experiment had little (if any) carry-over effect on this ecosystem property.

An increased invasion resistance in species-diverse grassland communities compared with less diverse communities has been found in studies where initially unsown species have been deliberately introduced (Fargione *et al.*, 2003), where unsown species have been removed over time (Knops *et al.*, 1999; Hector *et al.*, 2001; Tracy and Sanderson, 2004b), when ingress by unsown species in the harvested biomass was determined (Picasso *et al.*, 2008), and in surveys of pastures across a wide geographic area (Tracy and Sanderson, 2004b). Similar results have also been obtained for annual crops (Hauggaard-Nielsen *et al.*, 2008). However, plant functional diversity effects on invasion resistance have not been extensively studied. In this context, it is of interest to note that the most common of the unsown species observed in the experiment belonged to a functional type (deep-rooted forb) that was not present in the seeding mixtures (Fargione *et al.*, 2003; Hooper *et al.*, 2005). Species with competitive traits may, however, play a key role (see below). The causal mechanism for the increased invasion resistance by species-diverse communities is likely to be a more complete utilization of environmental resources (Hooper *et al.*, 2005). Differences in temporal growth patterns between the grasses and legumes currently used is, for example, a generally accepted agronomic fact (see Table S2 in Supplementary data available online).

Of the monocultures, the cocksfoot and two of the white clover (N and BC) communities were least susceptible to invasion, while all red clover monocultures were the most invaded. The reason why the BN white clover monoculture did not conform to the pattern might be that it had a lower growth rate at the end of the growing season, leaving more 'space' for unsown species to enter the sward at this time of the year. Since an efficient herbicide treatment was applied to this experiment shortly after sowing, the ingress of unsown species reflected characteristics of the mature community. Mwangi *et al.* (2007) found that grasses exerted a strong negative effect on invaders, whilst legumes had a positive effect, which to some extent agrees with the results from this study, since red clover resisted invasion poorly. Crawley *et al.* (1999), reporting a 7-year experiment with four levels of species richness where unsown species were allowed to accumulate in the plots, concluded that species identity mattered more than species richness in determining both the number of invading species and the total biomass of the invaders. Legumes play a vital role in the nitrogen economy of grassland communities (Spehn *et al.*, 2002; Temperton *et al.*, 2007), and it has been found that community invasibility is positively correlated with nitrate availability in the soil (Dukes, 2001). However, since all white clover monocultures contained substantially lower amounts of unsown species than the red clover monocultures, this cannot be the only mechanism that facilitated invasion of unsown species into the red clover plots. The fact that red clover plants died throughout the experimental period, leaving large gaps that could provide entry points for invading species, was probably just as important (see, for example, Milbau *et al.*, 2003, 2005).



One trait common to cocksfoot and white clover is the ability to spread laterally, cocksfoot both above- (Lorentzen *et al.*, 2008) and below ground (Personeni and Loiseau, 2004), and white clover above ground (Frame *et al.*, 1998). This trait would result in a more complete resource utilization, thus preventing the establishment of unsown species. The present monoculture results (Table 4) are in accordance with the suggestion made by Richardson and Pyšek (2006) that, in future, more emphasis should be placed on species identities and their characteristics in the search for an increased understanding of community invasibility, but in this study the species mixture results showed an even stronger effect.

It is concluded that increased species diversity within a forage crop provided clear and persistent yield benefits. This biodiversity effect could be profitably exploited by grassland farmers, and has the potential to reduce requirements for fertilizer inputs. Using broad genetic-base composites of the legumes was of minor importance in terms of yield at this site, although the use of broad genetic-base populations containing site-adapted germplasm may have produced a different result. Invasion by unsown species was generally reduced by increased species diversity of the sward. However, the strong effect of individual species identity suggests that traits conferring a more complete resource utilization also contributed to control invasibility.

#### SUPPLEMENTARY DATA

Supplementary data are available online at [www.aob.oxfordjournals.org](http://www.aob.oxfordjournals.org) and consist of the following tables. Table S1: estimates of model coefficients for sown dry matter yield ( $\text{t ha}^{-1}$ ) for each of three harvest years. Table S2: estimates of coefficients for sown dry matter yield ( $\text{t ha}^{-1}$ ) for each of the three harvests within each of the three harvest years. Table S3: estimates of model coefficients of log (unsown species %) for each of the three harvest years.

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